

Research about the Removal Effect of the Constructed Wetland System on the Escherichia Coli K12

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Abstract—This paper analyzed the related methods to remove the biological indicator *Escherichia coli* K12 (*E. coli* K12) in the constructed wetland system (CWs), which included filtration and absorption of filling, effect of the plant root system and microorganism phagocytosis and so on. It indicated that *E. coli* K12 was mainly removed through filtration and absorption of filling, about 80% was removed by this process. Plants and microorganisms also contributed to the removal of *E. coli* K12. The removal load of *E. coli* K12 increased with the addition of pollution load. When hydraulic retention time (HRT) was 4d, there was a better linear relation ($R^2 = 0.9998$) between the removal load of *E. coli* K12 and pollution load. Meanwhile, it had a close relation between inactivation ratio of *E. coli* K12 and hydraulic load, the lower hydraulic load was, the higher inactivation ratio was.

Keywords—constructed wetland system (CWs); *Escherichia coli*; Inactivation ratio; HRT; Pollution load.

I. INTRODUCTION

The pathogenic bacteria in sewage have seriously threatened public health, it is one of the most serious public health problems against human's progress [1, 2]. People have paid close attention to the method about how to detect and sterilize the pathogenic bacteria in domestic wastewater effectively. *E. coli* is regarded as the most sensitive indicator of fecal contamination in the natural environment of water, soils and plants [3, 4, 5]. It has been the indicator bacteria for the Chinese drinking water standard all the time. At present, chlorine, chlorine dioxide, ozone, ultraviolet ray etc. are common methods for water sterilization. Recently, the constructed wetland system (CWs) has drawn great attention because of its prominent advantages, such as robust, low external energy requirements, easy to operate and maintain [6, 7]. It has been reported that in some CWs, the removal rate of *E. coli* and *Streptococcus caesium* reached 90% and more than 80%, respectively [8]. Song's [9] research indicated that the CWs could effectively remove 99.9% *Escherichia coli*. Other researches also showed that the CWs had good performance in the removal of pathogenic bacteria [10, 11]. However, most of them only studied single factor.

The study comprehensively studied effect of the fillings, hydraulic load, hydraulic retention time (HRT),

microorganism and plant etc. on the removal of *E. coli* K12, explored the optimal design parameters for the CWs to remove the pathogenic bacteria. Meanwhile, it initially studied the methods about removal of the pathogenic bacteria in the CWs, provided computing technique and design consideration for the large-scale management of rural domestic wastewater.

II. MATERIAL AND METHOD

A. Experimental facility

Four groups of testing apparatuses were used to simulate horizontal plug flow subsurface constructed wetland. The sewage flowed into the bed body from one side, then flowed through the filling bed horizontally after homogeneous water distribution, and outflowed from the bottom of the system. The whole cell body was made of plastic to avoid infiltration. Each wetland cell was $0.9\text{m} \times 0.5\text{m} \times 0.5\text{m}$, the bottom gradient was 1%. The system was divided into suction stage, processing stage, discharge stage.

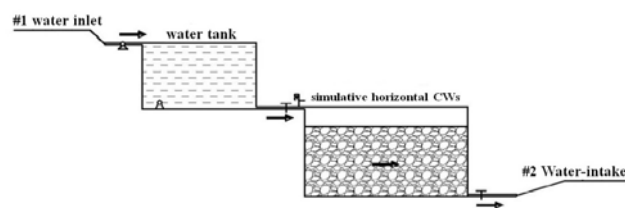


Figure 1. Diagrammatic sketch of experimental device

The water inlet of the testing apparatus was set at 0.1m over the system, the sewage was sent into the system from the inlet pipe. Globe valve was used to regulate water content for each testing apparatus. Overflow weir was set in the end to distribute water homogeneously.

The processing area was $0.9\text{m} \times 0.5\text{m} \times 0.45\text{m}$, and the bottom gradient was 1%. The fillings were consist of two parts: the super stratum was a mix matrix of coarse sand and soil (2:1) with a thickness of 25cm, the bottom was gravel, whose grain size is around 20mm, with a thickness of 15cm. The two parts were separated by geotechnical cloth. The planting density was 10 plants per square meter.

The water inlet was set at the bottom of the testing apparatus, 3cm over the ground. To avoid blocking, the water inlet was parceled with gauze and gravel was paved.

B. Experiment materials

In order to accomplish the research successfully, the necessary materials were as following: medium sand, whose grain size is between 0.35 and 0.5mm; gravel; calamus and clinopodiummurticifolium. The sewage used in the experiment was artificially made up, and E. coli K12 was bought from Shanghai Industrial Microbiology Institute.

C. Experimental contents

The designed running time was six months, except for the debug time, which would not exceed 15d. The experiment has been divided into four groups, and the detailed design was listed in Table 1.

TABLE I. ANALOG DEVICE TYPE

Group	Fillings	Others
W1	medium sand, gravel, soil	E. coli K12
W2	W21 medium sand, gravel, soil	calamus,E. coli K12
	W22 medium sand, gravel, soil	clinopodiummurticifolium, E. coli K12
Control group	no filling	E. coli K12

Firstly, in order to explore the influence of matrix filling, plant and activated sludge on the removal of pathogenic bacteria, we compared the control group with W1, W3 with W1, respectively. And then, at the condition of 28°C and 40cm water depth, the removal effect and load in the CWs were studied when the HRT was 2, 3 and 4d as well as the pollution load was 10^3 , 10^4 and 10^5 cfu/ml, respectively.

D. Detection method

Plate count was adopted to measure the E. coli K12 in the laboratory test. Microorganicinactivation ratio was used to evaluate the sterilizing effect in the CWs, the inactivation ratio was calculated as follows:

$$\text{Inactivation ratio} = \frac{N_0 - N}{N_0} \times 100\%$$

Where N_0 was the number of microorganism in the influent water, N was the left number of microorganism in the same amount of water after the treatment of the CWs.

III. RESULT AND DISCUSSION

A. Removal effect of filling on the E. coli K12

The results of natural inactivation of E. coli K12 and the filtration and absorption of the matrix fillings were shown in Fig.2.

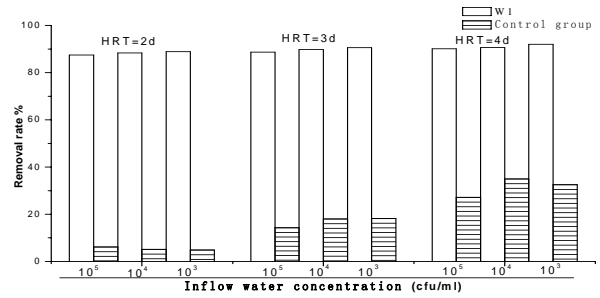


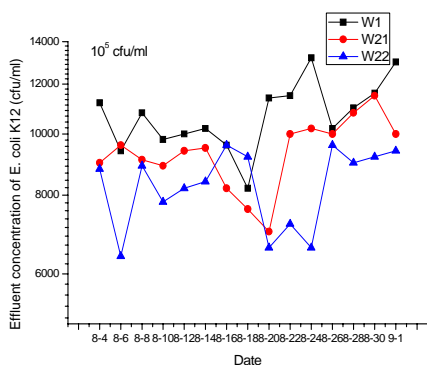
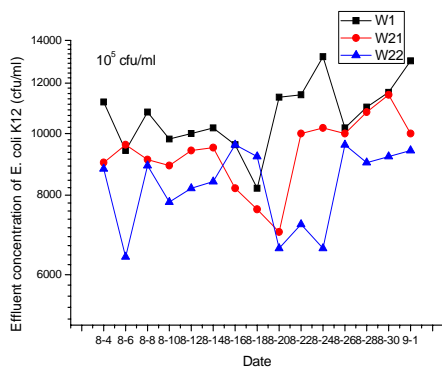
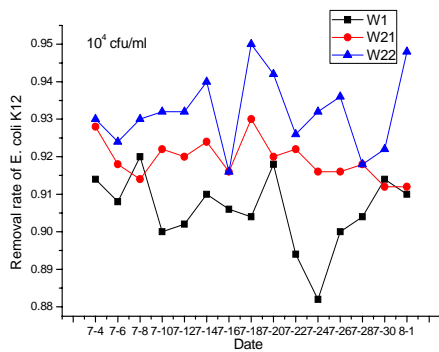
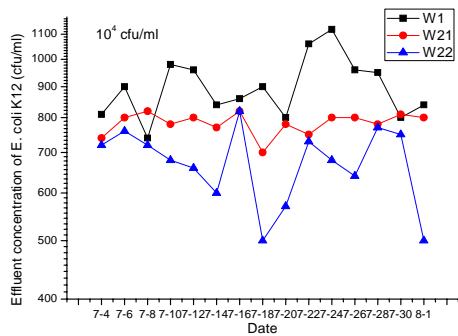
Figure 2. Removal rate of filtration and absorption of the matrix filling and natural inactivation of E. coli K12 under different HRT and inflow water concentration.

The result revealed that, the inactivation ratio of filling bed on the E. coli K12 increased with the increase of HRT when the influent water concentration ranged between 1×10^3 and 1×10^5 cfu/ml. When the HRT was 2d, the maximum removal rate of natural inactivation of E. coli K12 was only 6.2% in the control group, however, the minimum removal rate was 87.5% in group W1. So, the minimum removal rate of the filtration and absorption of the matrix fillings was 81.3%. When the HRT was 3d and 4d, the average removal rate of the filtration and absorption of the matrix fillings was 72.87% and 59.37%, respectively, the average removal rate of natural inactivation of E. coli K12 was 16.83% and 31.57%, respectively.

It indicated that the filtration and absorption of the matrix fillings was the main reason of the inactivation of E. coli K12. Meanwhile, at the same HRT, with the increase of pollution load in influent water, natural inactivation accounted for more in the whole removal rate of E. coli K12, while the filtration and absorption of the matrix fillings accounted for less. At the same pollution load, with the increase of HRT, the removal rate of natural inactivation of E. coli increased, while the filtration and absorption of the matrix fillings decreased, which could be due to the desorption along the wetland bed.

B. Removal effect of plant on the E. coli K12

Group W1, W31 and W32 were used for this experiment. The initial concentration of E. coli K12 was controlled around 10^4 cfu/ml, and then added up to around 10^5 cfu/ml half month later. The change of E. coli K12 in the CWs was shown in Fig.3.



As we could see in Fig.3, plant could help to promote the effluent quality of subsurface wetland to a certain degree, and, calamus performed better than clinopodium muticifolium. Under different initial concentration of *E. coli* K12, the average inactivation ratio of calamus on *E. coli* K12 was 93.19% and 91.60%, and the clinopodium muticifolium's was 91.92% and 90.35%, while the W1's was only 90.57% and 88.92%. That is because calamus has a more developed root, which could provide better breeding habitats for microorganisms. However, the effluent quality of clinopodium muticifolium was more stable.

C. Removal effect of pollution load on the *E. coli* K12

At the condition of 28°C and 40cm water depth, the removal effect and load of *E. coli* K12 in the CWs were studied when the HRT was 2, 3 and 4d as well as the pollution load was 10³, 10⁴ and 10⁵cfu/ml, respectively. The removal effect of *E. coli* K12 in the CWs under different HRT and pollution load was shown in Fig.4.

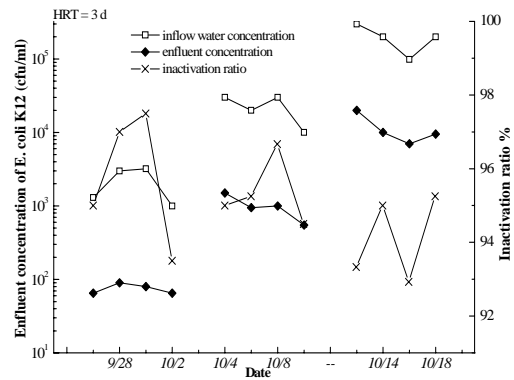
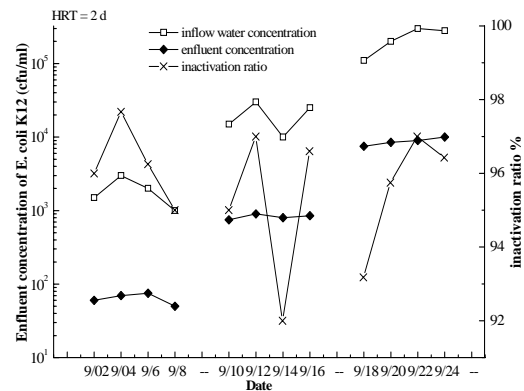


Figure 3. Removal effect of plant on the *E. coli* K12 under different initial concentrations of *E. coli* K12

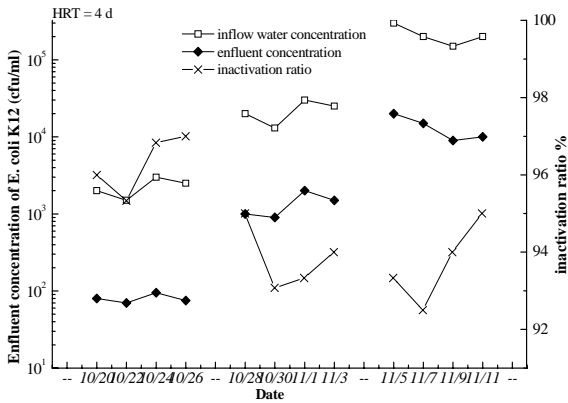


Figure 4. Removal effect of pollution load on the E. coli K12 under different HRT.

As we could see in Fig.4, at the same HRT, with the increase of pollution load, the effluent concentration increased, the removal rate of E. coli K12 decreased. At the same pollution load, with the increase of HRT, the effluent concentration decreased, and the removal rate of E. coli K12 increased. Moreover, when HRT was 4d, there was a better linear relation ($R^2 = 0.9998$) between the removal load of E. coli K12 and the inflow water load, and, with the increase of pollution load, it performed a corresponding better treatment effect.

D. Removal effect of hydraulic load on the E. coli K12

At the condition of 28°C and 40cm water depth, when the HRT was 3d, the removal effect of the CWs on the E. coli K12 under different hydraulic load was shown in Fig.6.

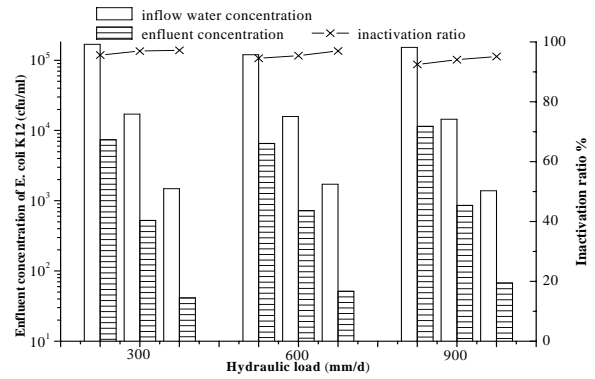


Figure 5. Removal effect of hydraulic load on the E. coli K12.

As we could see in Fig.5, the inactivation ratio of the CWs on the E. coli K12 had a close relation with hydraulic load. When the hydraulic load was 300, 600 and 900mm/d, the average inactivation rate was 96.58%, 95.65% and 93.90%, respectively. So, at the same pollution load, the higher the hydraulic load was, the worse the effluent quality. This is because that the higher hydraulic load would lead to shorter HRT, so the E. coli K12 would be brought out from the CWs before it could be adsorbed and filtrated, which would be bad for the purification of pollutant and reduce inactivation effect of pathogenic bacteria.

IV. CONCLUSION

1. Filtration and absorption of the matrix fillings was the main reason of the inactivation of E. coli K12. Meanwhile, at the same HRT, with the increase of pollution load in influent water, natural inactivation accounted for more in the whole removal rate of E. coli K12. At the same pollution load, with the increase of HRT, the removal rate of natural inactivation of E. coli increased, while the filtration and absorption of the matrix fillings decreased.

2. Plant could help to promote the effluent quality of subsurface wetland to a certain degree, and calamus performed better than clinopodium murticifolium.

3. At the same HRT, the effluent concentration increased and the removal rate of E. coli K12 decreased with the increase of pollution load. At the same pollution load, the effluent concentration decreased and the removal rate of E. coli K12 increased with the increase of HRT. Moreover, when HRT was 4d, there was a better linear relation ($R^2 = 0.9998$) between the removal load of E. coli K12 and the inflow water load, and, it performed a corresponding better treatment effect with the increase of pollution load.

4. The inactivation ratio of the CWs on the E. coli K12 had a close relation with hydraulic load. At the same pollution load, the higher the hydraulic load was, the worse the effluent quality. When the hydraulic load was 300, 600 and 900mm/d, the average inactivation rate was 96.58%, 95.65% and 93.90%, respectively.

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