

Enhanced Degradation of di-n-octyl phthalate (DOP) by Phenyllobacterium ESF-17 isolated from waste water treatment plant

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Abstract. In order to enhance the degradation of Di-n-octyl phthalate (DOP). One bacterial strain was isolated from activated sludge of waste water treatment plant(WWTP) using Di-n-octyl phthalate (DOP) as the sole source of carbon and energy. According to the phylogeny of 16S rDNA sequence, the bacterial strain ESF-17 was identified as Phenyllobacterium sp. Biodegradation of DOP by Phenyllobacterium ESF-17 was investigated. Results showed that the optimum pH and temperature for DOP degradation by Phenyllobacterium ESF-17 were 8.0 and 25°C. The strain exhibited higher degradation efficiencies and growth rates in alkalinity than in acidity. The strain could degrade approximately 93% of 400mg/L DOP shorter alkyl chain PAEs, whereas the degradation efficiency of longer alkyl chain PAEs were relatively poor. The results suggest that Phenyllobacterium ESF-17 may represent a promising application for DOP bioremediation.

Introduction

Di-n-octyl phthalate (DOP), as a prior pollutant for control, has been widely detected in environment. DOP is the member of phthalic acid esters (PAEs), which are widely used as plasticizers, building materials, additives in plastics and cosmetics production. [1] DOP is one of most common used PAEs. Three of the phthalic acid esters, namely, di-methyl phthalate (DMP), di-n-butyl phthalate (DBP) and di-n-octyl phthalate (DOP) have been listed as priority pollutants by China National Environmental Monitoring Center and the US Environmental Protection Agency [2] PAEs have received increasing attention in recent years due to their widespread produce, use and disposal, as a result they are also ubiquitous in environments. Moreover, These compounds are concerning because they have been shown to interfere with the reproductive system of human and animal. Phthalate esters (PAEs) are a prominent group of environmental pollutants and endocrine-disrupting compounds in many environmental. In addition, DOP can be taken up by crops and thus enter the food supply chain system, which may harm aquatic organisms and human health[3,4,5].DOP in one member of PAEs.

In recent years, many researches have been conducted to look for better way to degrade DOP including physical and chemical methods. Previous studies have revealed that DOP and DBP can be removed by natural processes in natural environments, such as hydrolysis, photo degradation and biodegradation [6,7,8]. Due to the low rate of chemical hydrolysis and photolysis of DOP, metabolic breakdown of this widespread pollutant by microorganisms is considered to be the major route. Several PAEs-degrading bacterial strains belonging to the genera *Sphingomonas*, *Pseudomonas*, *Rhodococcus*, *Microbacterium* and *Gordonia* have been isolated from different environments, such as active sludge.[7,8]

In the present paper, a DOP-degrading bacterium was isolated from active sludge and identified by 16S rDNA sequence. The biodegradation kinetics and different environmental factors affecting this process were investigated. The result from this study is expected to improve current understanding of the bioremediation of DOP and find more high effective DOP-degrading strains.

Materials and methods

Reagents and chemicals

DOP (99.5% purity) for the experiment was purchased from Chengdu Kelong Chemical Reagent Co., Ltd., All the chemical reagents were of analytical grade and all solvents (Ethyl acetate and methanol) were of HPLC grade purchased from Tianjing kemiu Reagent Co., Ltd.

The MM contained (1L): $\text{MgSO}_4 \cdot 7\text{H}_2\text{O}$ 0.5 g, K_2HPO_4 1.70 g, $\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$ 0.05 g, and NaNO_3 0.5 g, $(\text{NH}_4)_2\text{SO}_4$ 1.0 g, Na_2MoO_4 0.0024 g, $\text{CaCl}_2 \cdot 2\text{H}_2\text{O}$ 0.04 g, FeCl_3 0.0018 g.

The nutrient broth (NB) for bacteria enrichment consisted of beef extract 3g, peptone 5 g, NaCl 5 g, pH 7.2. Nutrient agar plates were made using NB supplemented with 2% agar.

Enrichment and isolation of DOP Strains

The enrichment procedure was according to Wu [9] with some modifications. Initially, 5.0 g of sludge was added to a 500-ml Erlenmeyer flask containing 200 ml of MM solution amended with concentration of 100 mg/l DOP. The suspension was incubated for 6 days in the dark at 25°C according to pre-experiment on a rotary shaker operated at 140 rpm. Subsequently, 2ml of the enrichment culture was serially transferred five times to fresh medium incubated under the same conditions. At the same time, in the process of transfer, containing a higher concentration of DOP 200–500 mg/l each time. Then the final enrichment was streaked onto MM agar plates supplemented with a mixture of DOP (500 mg/l) and incubated 1 week at 25°C. Presumptive colonies were picked on the basis of differences in colony morphology and coloration and re-streaked onto MM agar plates amended with DOP. The bacterial isolates were further purified by streaking on LB Nutrient Agar plates and then re-streaked onto MM agar plates with and without DOP to confirm their degradation abilities. Isolates can grow in the presence of DOP but not in their absence were selected for further study.

Amplification of 16S rDNA

Extraction kit (Sangon Corporation, Shanghai, China) was used for the extraction of bacterial genomic DNA according to the manufacturer's instructions. Further identification was performed by 16S rDNA gene sequencing, and then about 1500 bp length of 16S rRNA was amplified through PCR by using the bacterial universal primer 27F (50-AGAGTTTGATCCTGGCTCAG-30) and 1492R (50-GGCTACCTTGTTACGACTT-30). PCR was performed (Bio-Rad USA) under the following conditions: preheated at 95 °C for 2 min; then denatured at 94 °C for 1 min, annealing at 56 °C for 1 min, extended at 72 °C for 3min for 30 cycles, last extended at 72°C for 8 min.

Sequence analysis of strain

Purified PCR product was directly sequenced. The sequence data of the closest relatives were retrieved from NCBI database and aligned with CLUSTALW with all parameters set at their default values. A phylogenetic tree was then constructed using the neighbor-joining method with MEGA 6.0 software. The trees were validated using bootstrap analysis performed with 1000 replicates.

Degradation experiments of *Phenylobacterium* ESF-17

The following environmental factors were assayed to investigate their effects on DOP degradation within 60h of cultivation at a 140 rpm shaking rate. Temperature (10, 15, 20, 25 and 30°C); Initial pH value (4.0, 5.0, 6.0, 7.0, 8.0, 9.0); Initial DOP concentration (100mg/l, 200 mg/l, 300 mg/l, 400 mg/l, 500 mg/l). PAEs (DBP, DOP, DEP, DMP, DEHP and DPP)

Analysis method

Concentration of DOP in the supernatant solution was performed using high performance liquid chromatography (HPLC) (Agilent 1200 series). The column temperature was 40°C. The volume of the injected samples was 40µl; Chromatography column was Inertsil ODS-2151-K. 6× 150 mm.

Results and discussion

Isolation and identification of the DOP-degrading bacterium

Following 35 days enrichment, several DOP-degrading strains were isolated from the activated sludge. One strain shown high biomass and high degradation efficiency was selected for further investigation. Phylogenetic of the 16S rRNA gene (Fig.1) revealed strain *Phenylobacterium* ESF-17 clustered with members of the genus *Pseudomonas*, and had a 100% sequence similarity with *Pseudomonas amygdali*.

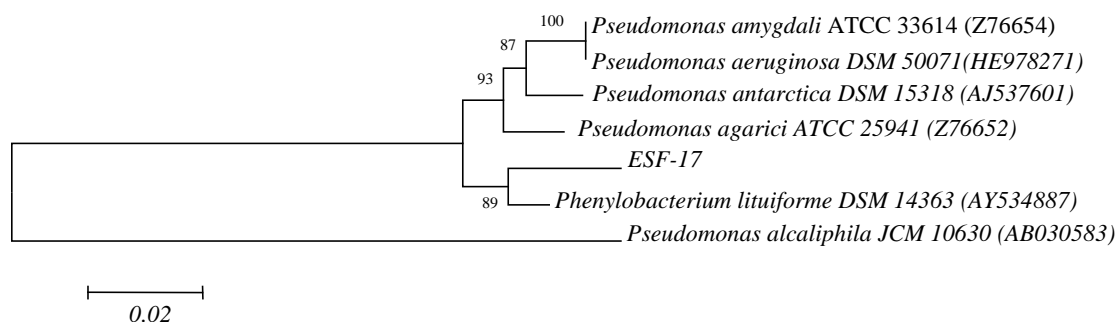


Fig. 1. Phylogenetic tree derived from 16S rRNA gene sequence of *Phenylobacterium* ESF-17 and sequences of related species. Distances were calculated using neighbor-joining method. Numbers at branch nodes are bootstrap values based on 1000 re-samplings. Scale bars represent 0.002 substitutions per site.

Effects of temperature on DOP biodegradation

The strain was cultivated at condition of 25°C, 500mg/L and pH8, at a 140 rpm shaking rate based on pre-experiment. The effects of temperature on the degradation of DOP in the culture medium were tested after incubation 60 h. The results showed that the optimal temperature for degradation was 25°C. The temperature are not consistent with previous reports [12][13]. There is no significant difference between 20°C and 25°C ($P < 0.05$).

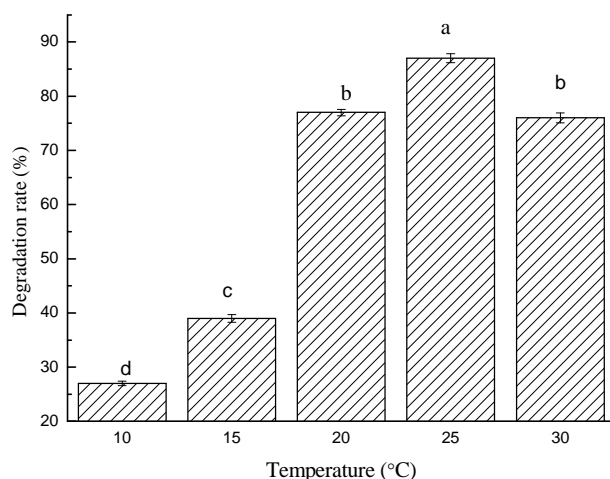


Fig. 2. Effect of temperature on degradation of DOP

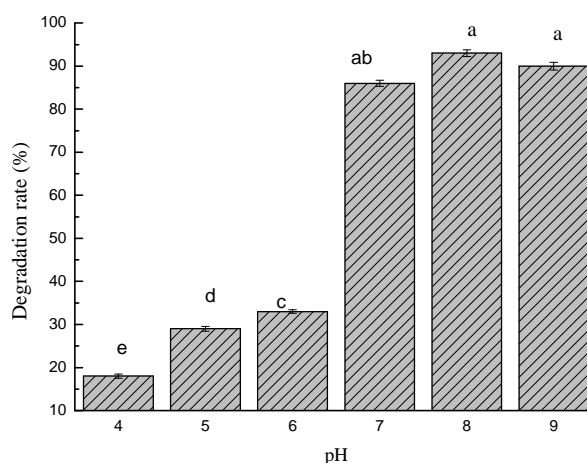


Fig. 3. Effect of pH and temperature on degradation of DOP

Effects of initial pH on DOP biodegradation

In order to investigate the effect of pH on strain degradation, different pH value was set. Fig.3. showed the Effects of pH (4.0–9.0) on DOP biodegradation at an initial concentration of 400 mg/L. We observed that the consortium exhibited higher degradation efficiencies and growth rates in alkalinity than in acidity. The highest DOP degradation rate was achieved at pH 8.0 (87.4%). The reported optimal pH values in degrading of various pollutants by other genus are ranging from 7.0

to 8.0[10] [11].The degradation rate of DOP decreased rapidly when pH decreased from 8.0 to 4.0. The results indicate that pH of 8.0 and 9.0 are optimal for DOP degradation and growth.

Effects of initial concentration on DOP biodegradation

In order to examine the effect of initial DOP concentrations on degrading efficiency by *Phenylobacterium* ESF-17. Experiment was conducted under DOP concentrations ranging 100, 200, 300,500 and 600 mg/L. Bacterial growth is concentration sensitive. As shown in Fig. 4 the DOP degradation rate decreased rapidly as concentration increased from 500 to 600mg/l. In general, the strain had a high degradation rate at the different concentration under 400 mg/L.

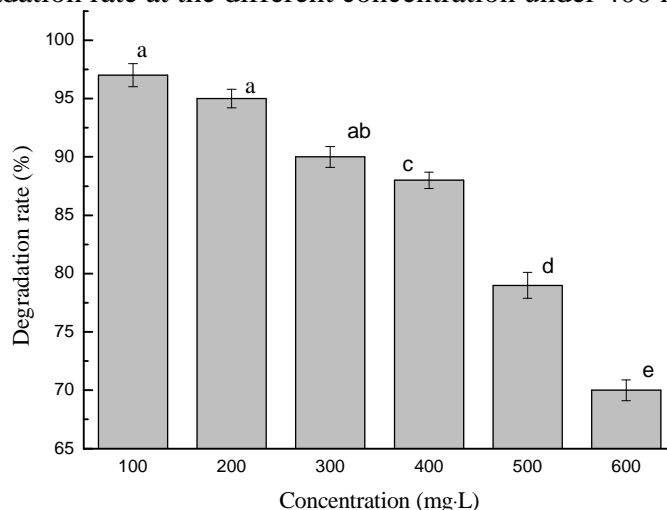


Fig. 4. Effect of initial concentration on degradation of DOP

Degradation of PAEs by bacteria

In order to investigate the degradation ability of the consortium to other commonly used PAEs in environment, the consortium was cultured in MSM supplemented with DBP, DOP, DEP, DMP, DEHP and DPP at 30 °C. Fig.5 presents the degradation rates for six PAEs and the biomass values of the consortium. The lowest biomass along with the lowest degradation rate was observed in DIOP. The consortium could utilize DBP, DOP, DEP, DMP, DEHP and DPP as the sole source of carbon and energy for growth. The degradation rates of DMP, DEP, DBP were higher than that of DOP, DEHP and DPP. The result indicated that shorter alkyl chain PAEs were degraded rapidly by consortium, while PAEs with longer alkyl chains were more difficult to be degraded.

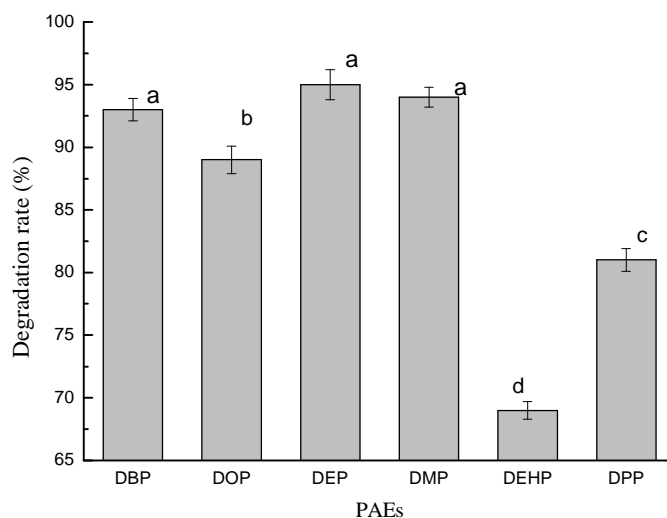


Fig.5. Degradation of other PAEs by isolate

Conclusions

A strain *Phenylobacterium* ESF-17 that can degrade DOP was isolated from activated sludge. Based on 16S rRNA sequence analysis, the strain was identified as *Pseudomonas amygdali*. This study investigated the optimal pH, initial concentration and temperature for DOP degradation in MM. The optimum biodegradation pH and temperature was 8.0 and 15°C, respectively. The optimal concentration was no more than 400mg/l. The study also showed under optimal pH and temperature condition, strain *Phenylobacterium* ESF-17 was capable of degrading DOP up to 94% in 6d incubation, which suggest that *Pseudomonas amygdali*. *Phenylobacterium* ESF-17 is a potential candidate for DOP degradation.

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