

Simulation of Spatio-temporal Variations of Methylmercury in Aha Reservoir, Guizhou, China by WASP

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Abstract—Two-dimensional, 8-segment mercury module of Aha Reservoir (surface area 4.15 km^2 ; volume $0.445 \times 10^9 \text{ m}^3$) was developed using the WASP 7 modeling system. The model calculated the concentration of methylmercury for the water column generally agreed with the measured values reported in literature for Aha Reservoir. Combined ArcGIS 9.3 and WASP 7, the spatial and temporal variations of methylmercury concentration were got in Aha Reservoir. The MeHg concentration was different for eight segments and increased for segments 1, 4, 5, 7 and 8, decreased for segments 3 and 6, and changed little for segment 2 from March to August in 2005.

Index Terms—Water quality; WASP 7; Mercury module; Methylmercury (MeHg); The Aha Reservoir; Spatio-temporal variation

I. INTRODUCTION

Mercury is a very toxic metal. Many studies have been dedicated to mercury cycle in the environment, the physical and biogeochemical transport and transformation of mercury within aquatic systems, and the chemical and biological mechanisms that effect the transformations of mercury in water [2][3][4][5].

Besides elemental mercury (quicksilver), the major forms of mercury in water are divalent mercury (which is bound to chloride, sulfide, or organic acids) and organic mercury, particularly methylmercury (MeHg). While all mercury species is very toxic, the organic mercury species rather than inorganic mercury can bioaccumulate because it is the only form of mercury that is better retained by organisms at various levels in the aquatic food chain [6]. MeHg accounts for the majority of the organic mercury species in freshwater systems. The most toxic form of mercury is MeHg. Even very low concentrations of MeHg in water, bioaccumulation through food web may cause high levels of mercury contamination in fish from in aquatic systems [7]. Aquatic food consumption is the primary source of MeHg exposure in humans. The toxic effects of MeHg are well known. In Japan, many thousands more people were affected and more than 1000 have died due to Minamata Disease which is MeHg poisoning that occurred in humans who ingested fish contaminated with MeHg [8][9]. So, MeHg is the species of greatest concern for both human health and ecosystems [4][10]. Some country established water quality criteria for MeHg to protect human health and aquatic life [11].

In this work, the distribution of the MeHg concentration of Aha Reservoir was studied. Lake Aha, is the drinking water

source of Guiyang City, China [12][13][14]. Its water supply capacity is 0.23 million tons per day. In Aha Reservoir, the concentration of MeHg exceeds the concentration in the other natural lake in the world [14]. From March to August 2005, it was measured and researched the concentration, transportation and fate of mercury species in lake water columns, lake tributaries water, lake sediment profiles and fish from Aha reservoir [12][13][14]. Nonetheless we still have an incomplete understanding of the spatio-temporal distribution of the concentration of MeHg in Aha Reservoir.

II. STUDY SITE

Lake Aha, which is located in the suburbs of Guiyang City, Guizhou Province, China, is a medium-sized artificial reservoir built in 1960. Lake Aha has a surface area of 4.15 km^2 , a volume of $0.445 \times 10^9 \text{ m}^3$ (a length of 517 km and a maximum width of 419 km), and mean and maximum depth of 13 and 24 m, respectively. Its mean water flux is $1.02 \times 10^9 \text{ m}^3 \cdot \text{a}^{-1}$, and the residence time of water is 0.44 a. It slightly remains thermally stratified in the summer and autumn [14][15].

Supplying water for Lake Aha, the main inflowing rivers are Youyu River, Baiyan River, Sha River, Lannigou River and Caichong River (Fig. 1). The only outflow river is Xiaoche River. Lake Aha was contaminated by the coal mines waste water and domestic sewage. After 1990's, domestic sewage and industrial waste water were the main pollution source in the Aha Reservoir for the mine waste water was controlled. The main pollution comes from the coal mines waste water carried by Youyu River, Baiyan River and Caichong River and the domestic sewage and industrial waste water from Sha River and Lannigou River [12][13][14].

The concentration of different mercury species in the water columns and sediment profile collected from two sampling sites of Aha Lake were investigated in 2005 [12][13][14]. The concentrations of total mercury, MeHg and reactive mercury are 7.1-47.9 ng/L, 0.16-2.5 ng/L and 0.25-2.43 ng/L respectively. Discernible seasonal variation of different mercury species were obtained during the sampling periods, with the concentrations in high flow period generally higher than those in low flow period [12][13][14]. However the spatial distribution of different mercury species was not get by two sampling sites. So, in this study, we studied the spatio-temporal distribution of the

MeHg concentration for it is the most important form of mercury that accumulates in aquatic food webs.

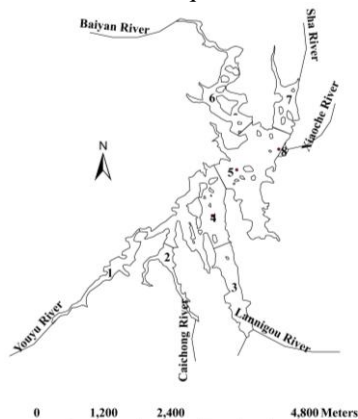


Fig.1 Configuration of mercury model segments with segment number and river in Aha Reservoir

III. MATERIALS AND METHODS

In this work, the Water Quality Analysis Simulation Program (WASP 7, USEPA) model was used for simulation of mercury transport and transformation processes in the Aha Reservoir. WASP is applicable to all kinds of aquatic systems. A subset of the WASP7 general toxic chemical module TOXI7 is used by the WASP7 mercury module MERC7 to simulate mercury cycling and transport through a water body. MERC7 simulates three mercury species, elemental mercury (Hg^0), divalent Hg ($Hg(II)$), and MeHg, as well as one to three solids types (i.e., silt, sand, biotic solids). MERC7 calculates mercury species and solids concentrations in the water column and sediments of each reach throughout the simulation period. With user-specified partition coefficients, $Hg(II)$ and MeHg are partitioned to suspended and benthic solids and to dissolved organic carbon (DOC). The model includes several transformation reactions of mercury species that are oxidation of Hg^0 in the water column, reduction and methylation of $Hg(II)$ and demethylation of MeHg in the water column and sediment layers. A detailed description of WASP 7 and the mercury module model is provided in WASP 6 manual and the overview of the Wasp7 Mercury Module which is available in the Web [16]. Earlier versions of WASP have been used to examine the effects of remedial actions on mercury speciation and transport in a lake system [11]. Carroll et al. (2000) focuses on program modifications that predict bank erosion rates and Hg bank concentrations related to longitudinal slope by using the US EPA RIVMOD, WASP5 and MERC4 numeric codes [10]. Henry et al. (1995) modified WASP 4 included remineralization to simulate release of Hg from settling particulates for understanding the fate of Hg in Onondaga Lake, NY [17]. These studies concluded that the WASP model could accurately simulate Hg fate for the sites studied under given condition, especially the WASP7 is an enhancement of the original WASP [16].

The targeted lake, Aha Reservoir, was conceptually divided into eight segments(1,2,3,4,5,6,7,8) to represent the water column and the benthic sediment (Fig.1). The environmental conditions, mercury transformation rates and

partition coefficient may differ in these eight segments. Available modeling parameters such as mercury concentrations, physicochemical, biological and hydrogeological data were collected from the literature that dealt with Aha Reservoir and the remaining parameters were estimated based on the literature[12][13][14][15]. The geometry information of these segments was obtained from Google Earth by ArcGIS 9.3. The WASP 7.3's Mercury model is used for simulating a 5-month period, from March to August 2005, a period where field measurements are available in the literature [12][13][14]. In this model, WASP 7.3 is set to calculate net flow transport across a segment interface, sediment bed volume statically and modeling time step automatically. The averaged velocity fields obtained were used to simulate transport and dispersion of mercury. Several constants and parameters contained in mercury reaction equations must be specified to implement the WASP7 mercury model. Calibrated constant parameters parameter is provided in table. 1.

TABLE.1 CALIBRATED CONSTANT PARAMETERS FOR WASP 7 MERCURY MODEL

Constant Parameter	Value
Elemental Mercury	
Partition Coefficient of Hg^0 to Silts and Fines (L/kg)	0
Log KDOC Partition Coefficient for Hg^0 (L/kg)	0
Reaction Yield Coefficient Elemental to Divalent Mercury (Mass Basis)	1
Henry's Law Constant for Elemental Mercury (atm-m ³ /mole)	0.0071
Divalent Mercury	
Partition Coefficient of $Hg(II)$ to Silts and Fines (L/kg)	500000
Partition Coefficient of $Hg(II)$ to Sands (L/kg)	1000
Log DOC Partition Coefficient for $Hg(II)$	4.5
Measured surface photoreduction rate constant for $Hg(II)$, 1/day	0.01
Reaction Yield Coef. for photoreduction of divalent mercury to elemental mercury	1
Quantum Yield for Dissolved Divalent Mercury	1
Quantum Yield for Sediment-Sorbed Divalent Mercury	1
Methylation Rate Multiplier for Dissolved Divalent Mercury	1
Methylation Rate Multiplier for Sediment-Sorbed Divalent Mercury	1
Reaction Yield Coefficient for Water Column Methylation (Mass Basis)	1.07
Reaction Yield Coefficient for Benthic Methylation (Mass Basis)	1.07
Methyl Mercury	
Partition Coefficient of MeHg to Silts and Fines (L/kg)	500000
Partition Coefficient of MeHg to Sands (L/kg)	1000
Log DOC Partition Coefficient for MeHg	5.5
Bacterial Demethylation Rate Multiplier for Dissolved MeHg	1
Bacterial Demethylation Rate Multiplier for Sediment Sorbed MeHg	1
Yield Coefficient for Bacterial Demethylation of MeHg to $Hg(II)$, g/g	0.93
Surface photo-demethylation rate constant, 1/day	0.05
Yield Coefficient for photoreduction of MeHg to Hg^0 , g/g	0.93
Quantum Yield for Dissolved Methylmercury	1
Quantum Yield for Sediment-Sorbed Methylmercury	1

IV. RESULTS AND DISCUSSION

The model was calibrated by adjusting the constants parameters in different segments. Generally, the observations agreed with simulation results for the concentration of MeHg. Only the difference between models and observations in segment 3 is large while those of other segments were small (Fig.2). An acceptable agreement was achieved.

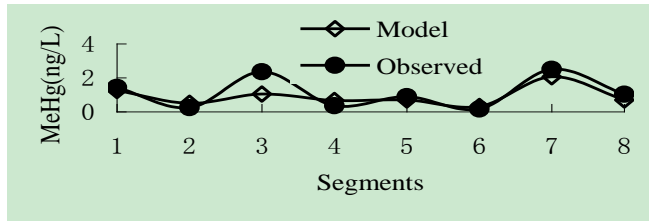


Fig. 2 Validation charts. Observed data (August, 2005) and model results

Coverage of Aha Reservoir and related database was developed using the ArcGIS 9.3 program. The database which has one field, SEGID, is used to align the mercury model predicted results with the correct segments. Then the spatial and temporal variations of the concentration of MeHg are analyzed.

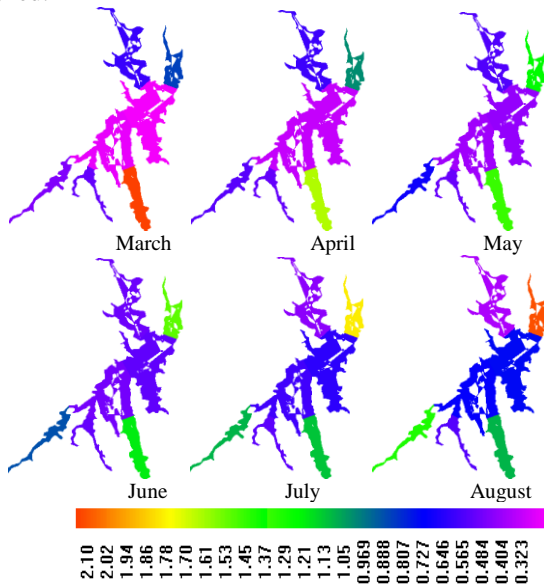


Fig.3 The spatial distribution of simulated concentration of MeHg (ng/L) in Aha Reservoir in 2005

The spatial variations of the concentration of MeHg in Aha Reservoir from March to August in 2005 were observed (Fig.3 and Fig.4). The concentration of MeHg was different in these eight segments from March to August. Except segments 4, 5 and 8, the difference of other segment was distinct. There was little difference in MeHg concentrations between segments 4, 5 and 8. In addition, the difference in the concentration of MeHg between segments 5 and 8 is smaller than that between segments 4 and 5. That the difference in the MeHg concentration among segments is obviously can be explained by the inflow. The segment 8 was connected with the only outflow river of Aha Reservoir, Xiaoche River and segment 5. So it is not affected directly by water quantity and quality of other inflow water and only by segment 5. This may be the reason of the difference of the MeHg concentration in

segments 5 and 8 is the least. In contrast, the segments 4 and 5 were influenced by different inflow rivers with different discharges and loads. This may lead to the difference of the MeHg concentration in segments 4 and 5 is bigger than that in segments 5 and 8.

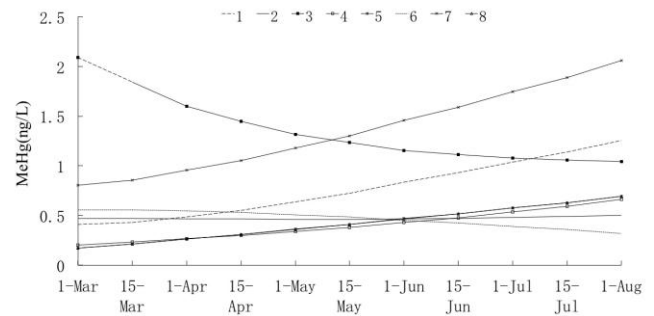


Fig.4 The temporal variation of simulated concentration of MeHg (ng/L) in Aha Reservoir from March to August, 2005

Significant temporal variation of the concentration of MeHg for each segment was observed (Fig. 4). The MeHg concentration increased for segments 1, 4, 5, 7 and 8, changed little for segment 2 and decreased for segments 3 and 6. The temporal variations of MeHg concentration in segments 1, 4, 5, 7 and 8 agreed well with the previous experimental results [12][13]. MeHg and other mercury species in these segments increased from the low flow period (March) to the high flow period (August) in Aha Reservoir. However, interestingly, for segment 3 and 6, our simulation result was contrast to the previous experimental results [12][13]. In general, the MeHg concentration is low in the high flow period for the water dilution. The input river of the lake Aha, Youyu River, Baiyan River and Caichong River all flow through the coal mines areas. The temporal variation of MeHg concentration in segments 1, 2 and 6 should be similar. Only considered the effect of river inflow, we can not understand the variation.

The observed spatial and temporal variations of the concentration of MeHg were related to several factors, such as the mercury processes included by WASP and parameters of hydraulic geometry, environment, transport, boundary, transformation and partition. So do that of the Aha Reservoir water body. Among the segments, one factor is different, the MeHg concentration will differ. In order to better understand the spatio-temporal variation of the MeHg concentration, further study is needed.

V. CONCLUSION

The MeHg concentrations in the water column were simulated by using WASP. The modelling results consist to the measured mercury concentrations reported in literature in the water compartment of the Aha Reservoir. The spatio-temporal distribution of the concentration of MeHg in Aha Reservoir was achieved by combining ArcGIS and WASP 7. The MeHg Concentration of the eight segments is different, except those of segments 4, 5 and 8 have some similar. In addition, from March to August in 2005, the MeHg Concentration increased for the segments 1, 4, 5, 7 and 8, decreased for segments 3 and 6, and changed little for segment 2. The cause for the spatio-temporal variation of the

MeHg concentration is very complex. It is related to all kinds of mercury processes in WASP and other parameters such as hydraulic geometry, transport, boundary, transformation and partition. For analyzing the reason of the variation, further study is needed.

ACKNOWLEDGMENT

The authors are grateful to Fulong Chen and Qinglong Fu of Xinjiang Institute of Ecology and Geography, Chinese Academy of Sciences (CAS) for their help in the experiments. The able technical assistance provided by ShuYong Mu (the Director), Li Zhao, Xin Wang, and Qianqian Du of the Center laboratory of Xinjiang Institute of Ecology and Geography, CAS is gratefully acknowledged. Financial assistance from National Natural Science Foundation of China (NSFC) (No. 41101040), "Western Light" Talents Training Program of CAS (XBBS201005, XBBS201010), Knowledge Innovation Program of Chinese Academy of Sciences (KZCX2-YW-335) and Program of 100 Distinguished Young Scientists of the Chinese Academy of Sciences is thankfully acknowledged.

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